

Research Article

Machine Learning Prediction of Cyanobacterial Toxin (Microcystin) Toxicodynamics in Humans

Stefan Altaner^{1#}, Sabrina Jaeger^{2#}, Regina Fotler¹, Ivan Zemskov³, Valentin Wittmann³, Falk Schreiber^{2,4} and Daniel R. Dietrich¹

¹Human and Environmental Toxicology, University of Konstanz, Konstanz, Germany; ²Life Science Informatics, University of Konstanz, Konstanz, Germany; ³Organic and Bioorganic Chemistry, University of Konstanz, Konstanz, Germany; ⁴Faculty of IT, Monash University, Melbourne, Australia

Abstract

Microcystins represent a family of cyclic peptides with approx. 250 congeners presumed to be harmful to human health due to their ability to inhibit ser/thr-proteinphosphatases (PPP), albeit all hazard and risk assessments are based on data of one MC-congener (MC-LR) only. Microcystin congener structural diversity is a challenge for the risk assessment of these toxins, especially as several different PPPs have to be included in the risk assessment. The inhibition of PPP1, PPP2A and PPP5 by 18 structurally different microcystins was determined and demonstrated microcystin congener-dependent inhibition activity and a lower susceptibility of PPP5 to inhibition than PPP1 and PPP2A. The data were employed to train a machine learning algorithm that allows prediction of PPP inhibition (toxicity) based on 2D chemical structure of microcystins. IC₅₀ values were classified into three toxicity classes, and three machine learning models were used to predict the toxicity class, resulting in 80-90% correct predictions.

1 Introduction

Harmful (toxic) cyanobacterial blooms have become an important concern with regard to drinking water quality and safety. Some prominent examples are a bloom that affected almost 1000 km of the Barwon-Darling River, New South Wales, Australia, in November and December 1991 (Bowling and Baker, 1996), the deaths of renal dialysis patients in 1996 in Caruaru, Brazil (Azevedo et al., 2002), or the more recent closing of the drinking water supply for the inhabitants of Toledo, Ohio, USA, resulting from recurrent Microcystis aeruginosa blooms in Lake Erie (Berry et al., 2017). The impression that cyanobacterial blooms in surface waters appear to be increasing with climate change is of current concern (Huisman et al., 2018). It is of importance in conjunction with toxic cyanobacterial blooms that several different toxins and congeners of a given toxin (e.g., microcystins (MC)) can co-occur in a given bloom (Dietrich and Hoeger, 2005). Also, toxin concentrations per cyanobacterial cell can change more than ten-fold within the short time span of a bloom event (Wood et al., 2011),

and higher temperature may provide a growth advantage for tox-in-producing species (Kleinteich et al., 2012).

MCs, which are produced by several cyanobacteria species, e.g., Microcystis spp., Dolichospermum spp. or Planktothrix spp., in water bodies worldwide (Preece et al., 2017), represent one of the toxin types most frequently associated with drinking water, food supplement and/or food contamination and have resulted in human health morbidity and mortality. Structurally, MC are cyclic heptapeptides consisting of common L-amino acids, but also uncommon and unique amino acids. Their general structure is cyclo([D-Ala1]-[L-X2]-[β-D-MeAsp3]-[L-Z4]-[Adda5]-[y-D-Glu6]-[Mdha7]). X and Z stand for variable L-amino acids, while β-D-MeAsp is erythro-β-D-methylaspartate, ADDA is (2S,3S,8S,9S,4E,6E)-3-amino-9-methoxy-2,6,8-trimethyl-10phenyl-4,6-decadienoic acid and Mdha is N-methyldehydroalanine. The variable positions, along with various (de)methylation sites (Fig. 1, Tab. S11), provide for currently 248 known MC congeners (Spoof and Catherine, 2017), albeit new MC congeners are continuously being discovered. However, contrary to a recent

Received April 3, 2019; Accepted June 27, 2019; Epub July 2, 2019; © The Authors, 2019.

ALTEX 37(1), 024-036. doi:10.14573/altex.1904031

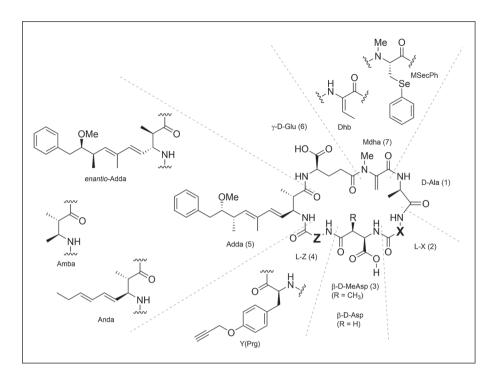
Correspondence: Daniel R. Dietrich, PhD Faculty of Biology, University of Konstanz, Universitätsstraße 10, 78457 Konstanz, Germany (daniel.dietrich@uni-konstanz.de)

This is an Open Access article distributed under the terms of the Creative Commons Attribution 4.0 International license (http://creativecommons.org/licenses/by/4.0/), which permits unrestricted use, distribution and reproduction in any medium, provided the original work is appropriately cited.

¹ doi:10.14573/altex.1904031s

[#] contributed equally





microcystins and the synthetic variants produced for this study
The dashed lines represent the single amino acids of the heptapeptide structure. Further details in Tab.
S1¹. Amba = (2S,3S)-3-amino-2-methylbutanoic acid, Anda = (2S,3S,4E,6E)-3-amino-2-methylnona-4,6-dienoic acid, Dhb = (E)-2-amino-2-butenoic acid, MSecPh = N-methyl-

Se-phenyl-L-selenocysteine, Prg =

proparqyl

Fig. 1: Consensus structure of

assertion (Huisman et al., 2018), the toxicity is known for only very few MC congeners.

The World Health Organization's provisional guideline value of 1 μ g/L for the risk assessment of MC in drinking water (WHO, 2017) is based entirely on the toxicological data of MC-LR and the assumption that MC-LR is the most toxic of the known MC congeners. This WHO guideline value is based on a 90-day toxicity study in mice (Fawell et al., 1999).

Toxicity can be split into two critical components: toxicokinetics (cellular uptake, distribution, metabolism and elimination) and toxicodynamics (the interaction with cellular molecules resulting in an observable adverse outcome) (Dellafiora et al., 2018; EFSA PPR Panel et al., 2018). Cellular uptake of MC is primarily governed by organic anion polypeptide (OATP) and thus by the type and level of OATP expression in a given cell as well as by the affinity and capacity of the respective OATP for transporting the different MC congeners (Fischer et al., 2010). Cellular export of MC (conjugates) is still under debate, as the involved exporters have so far not been unambiguously determined. However, a comparison of rodents with humans demonstrated that rodents are poor surrogates for humans with regard to the type of OATP expressed in the various tissues and the affinity and capacity of expressed OATPs for specific MC congener transport (Feurstein et al., 2011). The observation that humans demonstrate major differences in OATP expression and thus susceptibility to MC (Fischer et al., 2010) only compounded the fact that current risk assessment premises, based on surrogate species and one single MC congener, i.e., the WHO guideline value, could severely underestimate the potential toxicities of MC due to their congener-specific kinetics.

Similarly, questions have arisen with regard to the toxicodynamics of MC congeners. MC are very potent inhibitors of the catalytic subunits of ser/thr-protein phosphatases (PPP), however MC congeners differ with regard to their PPP inhibition capability (Hastie et al., 2005; Mackintosh et al., 1990; Hoeger et al., 2007). The family of PPPs in humans has seven members (PPP1, PPP2A, PPP2B (Calcineurin), PPP4, PPP5, PPP6 and PPP7), whose catalytic subunits are structurally similar (Shi, 2009). They display protein sequence homology of up to 65% (checked with Clustal Omega) and have defined substrate specificities and therefore differing functions (Pereira et al., 2011). Most of the PPPs are expressed ubiquitously, albeit at different levels in the various organs, while PPP7 is specific to retina and brain (Cohen, 2004). Inhibition of PPPs by MC occurs prima facie via reversible, followed by covalent, binding of MC to the catalytic subunit of the respective PPP (Mackintosh et al., 1990).

Dysregulated phospho-protein homeostasis subsequent to PPP inhibition, involving hyperphosphorylation of numerous phosphate-regulated enzymes and the deregulation of fundamental cellular processes, e.g., disruption of the cytoskeleton, thus represents the toxicodynamic process. As PPP differ in their susceptibility to inhibition by MC (Hoeger et al., 2007) and congeners differ in their capacity to inhibit specific PPP (Hoeger et al., 2007), the observed toxicity manifested in the respective organs (liver, brain, kidney) is not only the result of MC toxicokinetics but also of toxicodynamics (Fischer et al., 2010). To date, research on MC toxicodynamics primarily has focused on the interaction of MC congeners with PPP1 and PPP2A, whereby only a few MC congeners (predominantly MC-LR, -RR, -LA, and -LF) have been tested (Hoeger et al., 2007; Garibo et al., 2014).



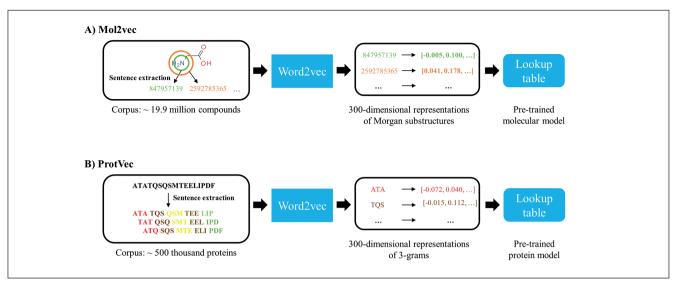


Fig. 2: Application of Word2vec to molecules and proteins

The procedure results in a lookup table, where 300-dimensional vectors can be extracted to have a numerical representation.

- A) For molecules, identifiers generated by Morgan fingerprint are considered as words, and their ordering as a sentence or molecule.
- B) For proteins, all possible 3-grams are considered as words, and by applying a sliding window over the protein sequence, three sentences are generated to represent a protein. Modified from Jaeger et al. (2018).

In view of the ever increasing number of identified MC congeners (Spoof and Catherine, 2017), yet lacking the ability to synthetize these in sufficient purity and amounts for *in vitro* or *in vivo* testing, an *in silico* approach using toxicodynamic data could provide a first step towards a better toxicity assessment of uncharacterized MCs with relevance for humans. However, as there are no sufficiently robust testing systems to address the toxicokinetic component yet, only the toxicodynamic component, i.e., PPP inhibition, was addressed in the present work.

Indeed, although a number of *in vitro* models have been put forward that allow studying the uptake of MC congeners (Fischer et al., 2010; Monks et al., 2007; Feurstein et al., 2009, 2010, 2011), studying the efflux from cells is much more difficult as intracellular MC would kill the cells before a proper efflux model could be established. Thus, despite recent advances in studying MC efflux using *in vitro* membrane vesicle approaches (Kaur et al., 2019), complete kinetic models encompassing influx and efflux kinetics of MCs have yet been impossible to establish. Accordingly, the aim of this study was to develop a comprehensive dataset of toxicodynamics, i.e., the PPP inhibitory capacities of a limited number of MC congeners.

These *in vitro* data were then used as a comparative basis driving an *in-silico* approach using machine learning (ML). Thus, the PPP inhibition capacity (toxicity) of 18 structurally diverse MC congeners was determined using ser/thr-PPP (PPP1, PPP2A and PPP5) in a colorimetric protein phosphatase inhibition assay. For this, a number of synthetic MC derivatives were generated according to previously published procedures (Zemskov et al., 2017; Fontanillo et al., 2016). Among these was a variant with modified stereochemistry at the Adda5 residue (i.e., the enan-

tiomer of Adda was used) ([enantio-Adda5]-MC-LF) as well as variants with simplified residues at the Adda5 position ([Anda5]-MC-LY(Prg) and [Amba5]-MC-LY(Prg)). The modified amino acids in these synthetic MCs in positions other than X2 and Z4 are indicated by adding the modification in square brackets before the name of the MC-XZ derivative. MC-LY(Prg) denotes variants with L-leucine in position X2 and O-propargylated L-tyrosine in position Z4 (Fig. 1, Tab. S1¹). Results were classified into three categories (toxic, less toxic, non-toxic) and the toxicity was predicted based on chemical structure via the ML approach described below.

Machine learning is widely used in the field of bioinformatics to predict bioactivity or molecular properties (e.g., solubility) of compounds, protein folding, etc. Despite recent advances, it is still difficult to employ ML in pharmacology or toxicology, as datasets are often smaller and more heterogeneous compared to datasets from other domains (Wu et al., 2018). Indeed, although ML has been employed for the prediction of cyanobacterial blooms based on satellite data (Chang et al., 2014) or the production of toxins based on environmental factors (Taranu et al., 2017), ML has so far not been used to predict the toxicity of MC congeners.

To encode molecules or proteins for ML, a fixed size numerical vector is needed (Wu et al., 2018). Approaches to encode molecules and proteins are Mol2vec (Jaeger et al., 2018) and Prot-Vec (Asgari and Mofrad, 2015), respectively, which are inspired by natural language processing. Both approaches are based on the Word2vec approach, which generates vector representation of words to capture semantic meaning (Mikolov et al., 2013). The vectors are obtained by training a deep neural network based on a database of text (so-called corpus) and results in a dense, high-di-



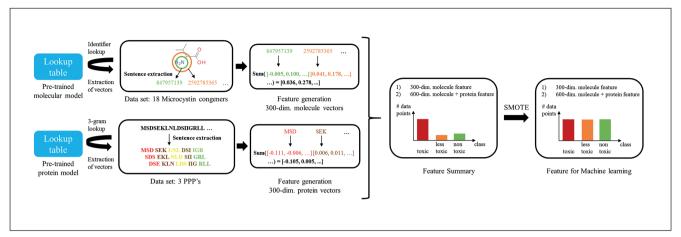


Fig. 3: Workflow for feature generation for toxicity classification

After downloading a pre-trained model for molecular structures and pre-training a protein model, vectors are extracted for substructures of 18 microcystin congeners and the 3-grams of the 3 PPPs. To represent a molecule or protein, the respective vectors are summed up. Then, either a 300-dimensional molecule vector is extracted, or combined with a protein vector, to build a 600-dimensional vector. As the dataset is highly imbalanced for the different classes, synthetic minority oversampling technique (SMOTE) was applied, to have the same number of compounds for each class. Modified from Jaeger et al. (2018).

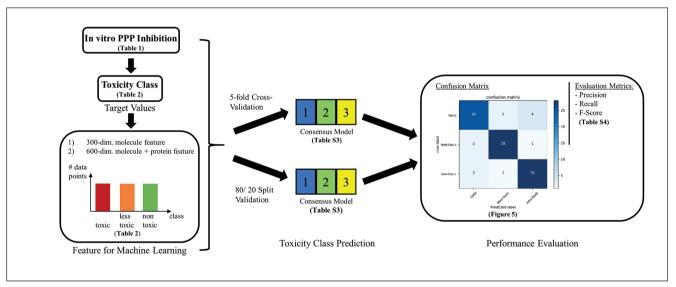


Fig. 4: After feature generation and pre-processing of the data, respective target values (toxicity class) are combined with the feature vector and used for a machine learning classification

Two different validation methods were used. For both validation methods, three machine learning models were set up, and majority voting was used for final prediction, building a so-called consensus model. Afterwards, performance was evaluated with a confusion matrix and evaluation metrics.

mensional representation of words. This procedure is a pre-training, resulting in a lookup table of words and vectors, which can be extracted later for various applications (e.g., ML).

To encode molecules with the Mol2vec approach, a large corpus of a database or collection of molecular structures has to be generated (Fig. 2A). The Morgan fingerprint is calculated for each molecule, but instead of hashing identifiers of substructures in a bit vector, identifiers (or words) are extracted and ordered to a sentence to represent a molecule. By this procedure, a molecular

lookup table of molecular substructures and vectors is generated, which is able to capture chemical relationships between substructures. To represent a new molecule, a Morgan fingerprint is generated, identifiers are extracted and looked up in the pre-trained model. Then, all substructure vectors are summed up to represent one molecule (Fig. 3), which results in a fixed size numerical representation of the molecule (Jaeger et al., 2018).

To encode a protein with the ProtVec approach (Fig. 2B), a large corpus of a database or collection of protein sequences has to be



Tab. 1: IC_{50} values of the tested MC congeners on rPPP1, hPPP2A and hPPP5

IC₅₀ were calculated after 5PL-nonlinear regression of at least 3 (hPPP2A) or 5 (rPPP1 and hPPP5) individual replicates using technical duplicates or triplicates. n.d. not determined (PPP2A no longer available, discontinued by manufacturer).

	rPPP1			hPPP2A			hPPP5		
Congener	IC ₅₀ (nM)	Cl ₉₅ (nM)	R ²	IC ₅₀ (nM)	Cl ₉₅ (nM)	R ²	IC ₅₀ (nM)	Cl ₉₅ (nM)	R ²
MC-RR	1.5	1.3 - 1.8	0.95	1.6	1.4 - 1.7	0.99	11.7	8.3 - 16.5	0.96
MC-LR	0.3	0.2 - 0.4	0.93	0.5	0.4 - 0.5	0.99	5.1	4.0 - 6.6	0.97
MC-YR	1.3	_	0.99	n.d.	n.d.	n.d.	5.1	4.3 - 6.1	0.99
MC-WR	1.2	1.0 - 1.5	0.94	1.0	0.8 - 1.1	0.97	5.6	4.2 - 7.6	0.97
MC-LA	1.9	1.4 - 2.7	0.86	0.7	0.5 - 0-9	0.93	6.1	4.3 - 8.7	0.96
MC-LY	0.8	0.7 - 0.9	0.99	n.d.	n.d.	n.d.	4.1	3.1 - 5.4	0.97
MC-LF	2.0	1.5 - 2.6	0.90	1.4	1.3 - 1.4	0.99	4.7	3.5 - 6.3	0.97
MC-LW	1.2	1.0 - 1.4	0.97	0.7	0.7 - 0.8	0.99	2.5	2.0 - 3.2	0.98
MC-HilR	0.6	0.5 - 0.8	0.99	n.d.	n.d.	n.d.	4.2	3.5 - 5.1	0.99
MC-HtyR	0.7	0.6 - 0.8	0.99	n.d.	n.d.	n.d.	4.7	3.6 - 6.0	0.96
[β-D-Asp3]-MC-RR	45.0	39.3 - 51.6	0.99	n.d.	n.d.	n.d.	167.1	131.8 - 211.8	0.97
[β -D-Asp3]-MC-LR	0.9	0.7 - 1.0	0.99	n.d.	n.d.	n.d.	10.2	8.3 - 12.5	0.99
[β -D-Asp3, Dhb7]-MC-RR	62.0	51.7 - 74.3	0.96	84.3	80.7 - 87.8	0.99	877.1	692.6 - 1111	0.97
MC-LY(Prg)	1.7	1.3 - 2.2	0.95	0.4	0.2 - 0.3	0.99	1.7	1.2 - 2.6	0.95
[MSecPh7]-MC-LY(Prg)	1.9	1.6 - 2.4	0.97	0.9	0.7 - 1.1	0.94	18.2	10.7 - 31.1	0.91
[enantio-Adda5]-MC-LF	_	_	_	_	_	_	_	_	_
[Amba5]-MC-LY(Prg)	520,817	449,800 - 603,048	0.98	2,135	1,991 - 2,291	0.99	54,063	37,431 - 78,087	0.95
[Anda5]-MC-LY(Prg)	1,724	1,434 - 2,072	0.98	n.d.	n.d.	n.d.	2,420	1,690 - 3,467	0.96

Tab. 2: Toxicity classes assigned to MC congeners

Classification is based on their PPP inhibitive capabilities (expressed as IC₅₀) for each of the three PPPs tested.

IC ₅₀ (nM)	Class	Description	Number of data points (original)	Number of data points after SMOTE
≤ 10	0	Toxic	31	31
> 10 ≤ 1000	1	Less-toxic	7	31
> 1000	2	Non-toxic	9	31
		Total	47	93

generated. Therefore, all possible n-grams of a protein sequence are generated by applying a sliding window over a protein sequence. This results in n-sentences to represent a protein. Those n-sentences are then used to generate the lookup table of protein n-grams. To represent a new protein, all n-grams are generated, looked up in the pre-trained model, and summed up to represent a new protein (Fig. 3), which results in a fixed size numerical representation of the protein (Jaeger et al., 2018; Asgari and Mofrad, 2015).

PPP inhibition data (IC₅₀ values) gained from the in vitro as-

says for the different MC congeners (Tab. 1) were then classified into three toxicity classes (Tab. 2, Fig. 3), and ML models were generated using the encoded MC congeners and PPP vectors (Fig. 2, 3). These were then trained with different features and classifiers to classify MC congeners into the toxicity classes, as shown in the ML flowchart (Fig. 4). The latter approach thus allowed us to predict the toxicity of MC congeners and to compare the predictions with the true findings from the *in vitro* PPP inhibition assays.



2 Materials and methods

Materials

Microcystins were obtained either from Enzo Life Sciences (MC-RR, -LR, -YR, -WR, -LA, -LY, -LF, -LW, -HilR, -HtyR, $[\beta$ -D-Asp3]-MC-RR and $[\beta$ -D-Asp3]-MC-LR) or were synthesized (MC-LY(Prg), [enantio-Adda5]-MC-LF, [Anda5]-MC-LY(Prg), [Amba5]-MC-LY(Prg), [MSecPh7]-MC-LY(Prg)) using previously published procedures (Zemskov et al., 2017; Fontanillo et al., 2016). [β -D-Asp3, Dhb7]-MC-RR was a gift from Judith Blom (University of Zurich, Switzerland).

Microcystins were dissolved in pure methanol to $100~\mu M$. Actual concentrations were determined using UV spectroscopy at 238 nm (using the extinction coefficient of MC-LR of 39800 mol L⁻¹ cm⁻¹), and stocks were stored at -20°C until used for serial dilutions. Stocks of [Anda5]-MC-LY(Prg) and [Amba5]-MC-LY(Prg) were quantified by dissolving weighed amounts in an appropriate volume of methanol, as photometric quantification was not possible due to lack of the characteristic absorption peak at 238 nm (missing ADDA chain).

rPPP1 (rabbit skeletal muscle) was obtained from New England Biolabs (P0754L, product discontinued). hPPP2A (human red blood cells) was from Promega (V6311, product discontinued). pET32a(+)-TrxA-6His-hPPP5 was generated by GenScript using the human PPP5 sequence (NCBI Accession NP_006238.1) with GenScript's codon optimization for *E. coli*.

Expression of 6xHis-hPPP5 in E. coli

pET32a(+)-TrxA-6His-hPPP5 was transformed into chemical competent BL21-CodonPlus(DE3)-RP $E.\ coli$ cells (Agilent, 230255) via heat-shock. After selection on ampicillin/chloramphenicol-agar, a single colony was picked and cultivated in 2 mL LB-ampicillin/chloramphenicol (amp/cam) medium for 6 h. Afterwards, the pre-culture was added to 500 mL of TB-amp/cam (Terrific Broth supplemented with ampicillin and chloramphenicol) medium and incubated over night at 37°C while shaking (220 rpm). The following day, OD₆₀₀ was measured to ensure growth in the exponential phase (OD₆₀₀ = ~4). Controls were performed using heat-shock treated BL21-CodonPlus(DE3)-RP $E.\ coli$ cells without plasmid.

Purification of 6xHis-hPPP5

Cultures of transformed BL21-CodonPlus(DE3)-RP *E. coli* cells were centrifuged at 5000×g for 10 min at 4°C. Pellets were washed with 10 mL STE buffer (10 mM Tris-HCl pH 8.0, 100 mM NaCl, 1 mM EDTA) prior to resuspension in 10 mL cold PBS buffer + 1% protease inhibitor cocktail (Sigma, P8849). Cell lysis was achieved using a Branson Sonifier 250 with five times 20 pulses. Samples were cooled on ice between each set of 20 pulses. Lysate was cleared from cell debris by centrifugation for 45 min at 18000×g at 4°C. The cleared lysate was incubated with 2.5 mL (50:50 slurry) equilibrated Ni-NTA-agarose beads (Biozym, 2631105) on an overhead-tumbler at 4°C overnight. After centrifugation at 4°C and 800×g for 5 min, the supernatant was discarded and beads were washed three times with 2.5 mL wash buffer (50 mM NaH₂PO₄, 300 mM NaCl, 60 mM imidazole). Elution was performed by incubating the beads for 10 min

at 4°C on an overhead-tumbler in 2 mL elution buffer (50 mM NaH₂PO₄, 300 mM NaCl, 100 mM imidazole). Samples were taken from each step for subsequent SDS-PAGE analysis. To avoid high imidazole concentrations in the final sample, buffer exchange via dialysis against storage buffer (20 mM TrisHCl pH 8, 100 mM NaCl) was performed using 2 L volumes for about 18 h with two buffer exchanges. 10% glycerol was added to the final samples before aliquoting and liquid nitrogen snap-freezing. All steps were either performed on ice or in a cooling room (4°C). Samples were stored at -80°C until further use. In addition, one analogous purification was performed using the heat-shock treated BL21-CodonPlus(DE3)-RP *E. coli* without plasmid ("empty expression") for control purposes.

SDS-PAGE analysis

SDS-PAGE analysis was performed using 10% SDS-gels in a Bio-Rad Mini-PROTEAN vertical electrophoresis system. 6 μL 6X SDS-sample buffer was added to 30 μL sample and the samples loaded into the gel pockets. Samples were not corrected for protein amount but for the volume of the original fraction in such a fashion that consistently 0.3% of the original fractions were used (filled up to 30 μL with MilliQ). Gels were run at 100 V for 60 min and stained using colloidal Coomassie solution at 4°C while shaking overnight. Images were taken using a desktop scanner.

Mass spectrometry

Mass spectrometry was employed to confirm the identity of the expressed TrxA-6×His-hPPP5. Samples were run on SDS-gels as described above and bands considered to contain TrxA-6×xHis-hPPP5 were cut out, divided into ~1 mm² squares and submitted to the Proteomics Core Facility of the University of Konstanz. Proteomic analyses of trypsin digested fragments were carried out with an LTQ Orbitrap Discovery (Thermo Fisher Scientific, Bremen Germany) coupled to an Eksigent 2D-nano HPLC (Eksigent, USA). Data were analyzed using Mascot software (Matrix Science).

Phosphatase activity assay

The phosphatase activity of the TrxA-6xHis-hPPP5 fraction (hPPP5) was assayed using a colorimetric phosphatase activity assay with para-nitrophenylphosphate (pNPP) as substrate and buffers according to Heresztyn and Nicholson (2001). Purified hPPP5 was tested undiluted and in 1:1 serial dilutions (total 11 concentrations, highest dilution 1:1024) using the enzyme diluent buffer (52 mM Tris pH 7, 2 mM MnCl₂, 0.5 mg/ml BSA, 1 mM DTT, 0.5 mM NaOAc, 123.5 µM EGTA) for dilutions. 20 µL of each dilution was pipetted into a well of a polystyrene flat bottom 96-well plate to which 200 µl of the substrate solution (62.5 mM Tris pH 8.1, 26 mM MgCl₂, 0.2 mM MnCl₂, 0.5 mg/ml BSA, 2 mM DTT, 1 mM NaOAc, 24 mM pNPP) was given. Testing was carried out in technical triplicates. Color development at 37°C and 405 nm was measured every 10 min over a period of 4 h. Linear regressions of each dilution over time were plotted and slopes obtained. Slopes obtained were compared to corresponding slopes of rPPP1 (protein amount and specific activity). Accordingly, this allowed calculation of the volume of the hPPP5



fraction corresponding to three U of rPPP1 (hPPP5: $0.83~U/\mu L$, rPPP1: $2.5~U/\mu L$). U is defined here as the amount of enzyme needed to dephosphorylate 1 nmol of pNPP in 1 min at 30°C.

Colorimetric protein phosphatase inhibition assay (cPPIA) The employed assay is based on previously published procedures (Heresztyn and Nicholson, 2001; Fischer et al., 2010). Serial dilutions of each MC congener were produced in MilliQ in LC-vials, whereby the most concentrated MC sample contained a maximum of 5% methanol. 20 µL of each dilution was pipetted into a polystyrene flat-bottom 96-well plate in triplicates. Three U of each phosphatase were diluted in 2120 µl enzyme dilution buffer and 20 µL of the solution was added to each well (corresponding to about 0.07 U per well). Enzyme solution (20 µL) in addition to 20 µL of MilliQ (three replicates) served as 100% activity control (no inhibition), while 20 µL enzyme dilution buffer lacking enzyme (three replicates) as well as 20 µL MilliQ served as background control (no enzyme activity). The plate was incubated at 37°C for 5 min to ensure interaction of microcystins with the PPP. 200 µL of substrate solution was pipetted into each well, and the plate was immediately read at 405 nm using an Infinite 200 Pro microplate reader (Tecan, Männedorf, Switzerland). The plate was then incubated at 37°C for 3 h before being measured again at 405 nm. PPP activity was calculated by subtracting the start value (0 h) from the end value (3 h) and compared to 100% activity. IC₅₀ were calculated using GraphPad Prism 5.0 software via a 5-PL non-linear regression with anchorage points and constraints between 100% and 0%. Replication: $n \ge 3$ for PPP2A and \geq 5 for PPP1 and PPP5, each in technical duplicates or triplicates. The analyses of [Amba5]-MC-LF and [Anda5]-MC-LF had 5 biological replicates but no technical triplicates due to shortage of pure testing material.

Data analyses, pre-processing, statistics and machine learning Data analyses were carried out using Microsoft Excel Professional Plus 2013, while GraphPad Prism 5 was used for statistics and data representation. Data pre-processing and machine learning were carried out using Python version 3.6.6².

Machine learning (ML)

The PPP inhibition capabilities of the different MC congeners, expressed as IC₅₀ values (Tab. 1), were used to train an ML model (Fig. 2-4). In order to allow for the adaption into the ML model, the two primary factors, i.e., MC congeners (molecules) and PPP (proteins), had to be transformed to numerical vectors (Fig. 2). To transform molecules to a vector, the Mol2vec approach described in Jaeger et al. (2018) was used with a pre-trained model³. To transform proteins to a vector, the ProtVec approach (Asgari and Mofrad, 2015) (Fig. 2) and a model trained on UniProt sequences were used as described in Jaeger et al. (2018). Subsequently, the models were applied to encode MC congeners and the PPPs (UniProt ID: PPP1 (P62136), PPP2A (P67775) and PPP5 (P53041)) as vectors.

PPP inhibition data of the different MC congeners (Tab. 1) were classified into three classes of toxicity (Tab. 2, Fig. 3, 4). The original dataset consisted of 47 data points of which 31 data points were classified as toxic, 7 were classified as less-toxic and 9 were classified as non-toxic. This classification was arbitrary, albeit the most toxic classification includes MC congeners with relevance to human intoxications and the WHO guideline (Berry et al., 2017; Dietrich and Hoeger, 2005; Azevedo et al., 2002; Bartram and Chorus, 1999). As this is a rather small dataset for ML and sufficient samples are crucial for ML performance. Synthetic Minority Oversampling Technique (SMOTE) was applied to mimic a balanced dataset and thereby increase prediction performance of minority classes (Chawla et al., 2002). SMOTE generates new, artificial data points for minority classes by variation of the feature vector representing original data points. SMOTE implementation in imbalanced learn (version 0.3.3) (Lemaître et al., 2017) was used with standard settings and a ratio of 1.0. This procedure resulted in 31 data points per class, i.e., 93 data points in total (Tab. 2).

Three different ML models were trained with different features and classifiers to classify MC congeners into the toxicity classes. Two models were trained with a random forest (RF) classifier implemented in scikit-learn (version 0.8.0) (Pedregosa et al., 2011) and the XGBoost implementation (version 0.19.1) of Gradient Boosting Machines (GBM) classifier (Friedman, 2001), respectively. As feature, vectorized structural data of the congeners (Mol2vec) was used. Additionally, one model was trained with a random forest classifier with molecular and protein data as feature (Tab. S3¹, Fig. 3, 4). To merge molecule and protein information into one vector, structural data of molecules was vectorized with Mol2vec and protein data was vectorized with ProtVec and then concatenated. Subsequently, hyper-parameters were tuned (Tab. S3¹) to derive the best model. The final model used majority voting of these three models (Mol2vec with RF, Mol2vec with GBM and Mol2vec + ProtVec with RF) to classify a compound (Fig. 4).

For training and evaluation of an ML algorithm, the dataset had to be split into a training and a test set. These two sets had to be strictly separated, because if data points from the training set were used in the test set, the evaluation would always result in high performance as the model would already know the data point. Two procedures were applied to split the dataset: 1) using 80% of the data points (75) for training and 20% of the data points (18) for testing the performance and 2) by five-fold cross-validation (Fig. 4): Applying five-fold cross validation results in four-folds of the dataset being used for training and one-fold of the dataset being used for testing. This procedure is repeated until every fold was the test set once, resulting in five ML models. This procedure has the advantage that a standard deviation of performance between the models can be calculated to get a better estimation of model performance and robustness (Tab. S4¹). For both procedures, data points are randomly assigned to the respective set or fold. For this reason, each procedure and ML was repeated 50

² https://www.python.org/

³ Mol2vec – an unsupervised machine learning approach to learn vector representations of molecular substructures. https://github.com/samoturk/mol2vec (accessed 08.08.2018)



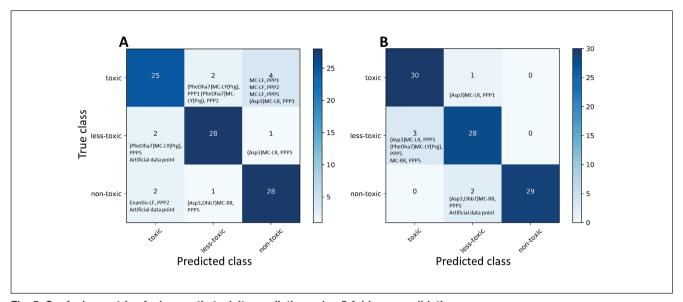


Fig. 5: Confusion matrix of microcystin toxicity prediction using 5-fold cross validation

Either the whole dataset was used for training (A), or the stereoisomer [enantio-Adda5]-MC-LF) was dropped for the training of the algorithm (B). The identity of all wrong classifications is given in the individual squares.

times to test whether performance was robust and independent of the random assignment of data points for training.

To finally estimate performance of the ML model, different performance metrics were employed (precision, recall, F-score, see Tab. S4 and Fig. S1¹). In addition, the confusion matrix (Fig. 5) was checked to estimate how many and which molecules were classified correctly or falsely. Performance metrics and confusion matrix were used as implemented in scikit-learn (Pedregosa et al., 2011).

3 Results

Full-length human PPP5 was bacterially expressed in BL21 Codon Plus *E. coli* with several attached tags: Thioredoxin A (TrxA) for solubility, 6-Histag for purification, S-Tag for a possible second purification, and a thrombin-site (TrxA-6His-S-PPP5). PPP5 identity was confirmed using mass spectrometry after purification using Ni-NTA beads (Tab. S2¹). To ensure that the observed activity was due to PPP5, bacteria without a plasmid were grown, purified and tested. These purified proteins did not show PPP activity (Fig. S2¹).

To develop a dataset of MC congener dependent toxicity (PPP inhibition activity) for the ML model, 18 different MC congeners were tested in three PPP (rPPP1, hPPP2A, hPPP5 expression). The 18 MC congeners spanned the known spectrum of hydrophobicity, had different molecular weights, and contained common as well as unusual modifications of the consensus structure (Fig. 1, Tab. S1¹). The *in vitro* PPP inhibition assays delivered well-fitted (R²) concentration-inhibition response curves (Tab. 1, Fig. S3-S¹); the derived IC₅₀ values were subsequently used for the ML approach (Fig. 2-4).

The comparison of IC₅₀ values obtained with MC congeners in the three PPPs demonstrated that of the ten MC congeners available for testing in both PPP1 and 2A, five MC congeners had comparable IC₅₀ values, while five MC congeners were more selective towards PPP2A, possibly suggesting a slightly higher sensitivity of PPP2A toward MC congeners (Tab. 1, Fig. S5¹). In contrast, several-fold higher concentrations of MC congeners were necessary to achieve 50% inhibition of PPP5 phosphatase activity. Notable exceptions to the latter were MC-LY(Prg) and [Amba5]-MC-LY(Prg), to which the PPP5 susceptibility to inhibition was comparable but lower than observed for PPP1.

The importance of structural differences with regard to binding to the catalytic subunit of PPPs was dramatically demonstrated with the comparison of MC-LF and the de novo synthetized [enantio-Adda5]-MC-LF. While MC-LF inhibited all three PPPs tested, the corresponding [enantio-Adda5]-MC-LF had no PPP inhibitory activity at all (Tab. 1). In contrast, other structurally similar derivatives, i.e., MC-LY(Prg) and [MSecPh7]-MC-LY(Prg), both having a propargyloxy residue at position 4 (Fig. 1) of the phenylalanine moiety, show only slightly reduced PPP inhibiting activity, if any, when compared to the parent MC-LF. However, if the Adda-residue is shortened to [Amba5]-MC-LY(Prg) or [Anda5]-MC-LY(Prg) (Fig. 1, Tab. S1¹), a marked reduction in PPP inhibiting capacity is found (Tab. 1). The latter observations suggest that structural changes of the amino acid Adda (enantiomeric configuration or shortened Adda-side chain) prohibited or reduced functional interaction with the catalytic subunits and thus inhibition of the PPPs. On the other hand, structural changes to the phenylalanine moiety at position 4 or the Mdha at position 7 had limited impact on PPP activity. Similarly, exchanging leucine in MC-LR and tyrosine in MC-YR for a homoisoleucine (MC-HilR) and a homoty-



rosine (MC-HtyR) at position 2 had no significant effect on PPP inhibition capacity (Tab. 1). However, changing the methylation of $\beta\text{-D-MeAsp}$ at position 3 of MC-RR to demethylated $[\beta\text{-D-Asp3}]\text{-MC-RR}$ and $[\beta\text{-D-Asp3},Dhb7]\text{-MC-RR}$ resulted in lesser PPP inhibition capacity, thus suggesting that structural changes involving L-amino acid residues at position 3 could have an impact on the inhibition of PPPs. When comparing the impact of structural changes of the Adda side-chain at position 5 with changes of $\beta\text{-D-MeAsp}$ at position 3, it appears that the former had a much more pronounced impact on the binding of MC congeners into the catalytic subunit and thus inhibition of PPPs.

The above IC₅₀ values were classified into toxicity classes (Tab. 2) and used for an ML approach. As the distribution of data points was not similar among classes, oversampling was applied, resulting in 31 data points per class, adding up to a total of 93 (Tab. 2, Fig. 3, 4). The resulting data was used for ML via two different approaches: using 80% of the data for training and 20% for testing of prediction (80/20) and by 5-fold cross validation (CV) with 50 repetitions (see Fig. 4, Tab. S4¹ and 3.9 for details).

Subsequently, all data points were used and majority voting of three ML models was employed to predict the toxicity class of every data point (Tab. S3¹). Both approaches of splitting data into training and test set performed well for toxicity class prediction with a precision above 0.8 and a recall and F-score mostly above 0.8 (Tab. S4¹). Since cross-validation is more suitable for small datasets (Beleites et al., 2013), we primarily focused on CV results, although 80/20 results are provided in Table S4¹ as well. CV predictions were compared with the true classification according to classified IC₅₀ values (Fig. 5A). 25 of the 31 toxic data points were predicted correctly, while two were wrongly predicted as less toxic and four as non-toxic.

Interestingly, the misclassified MC congeners primarily involved MC-LF variants. For example, the stereoisomer of MC-LF ([enantio-Adda5]-MC-LF) was classified as toxic in PPP2A although it is not toxic. The latter resulted most likely from the fact that the training vector generation was not trained for chirality of the molecules in question. Indeed, the vector generated for [enantio-Adda5]-MC-LF and MC-LF would be identical, albeit the values entered into the ML algorithm would read "non-toxic" and "toxic" and thus result in wrong classifications by ML.

To test this, the same approach was chosen but the [enantio-Adda5]-MC-LF-variant was omitted from the training and test sets of the ML algorithm (Fig. 5B). In consequence, both natural MC-LF and propargylated MC-LF variants were classified correctly. Moreover, [β-D-Asp3]-MC-LR (in PPP1) was now moved from originally being wrongly predicted as non-toxic to being predicted as less toxic, despite its true affiliation to the "toxic" class. Similarly, the prediction of [β-D-Asp3]-MC-LR (in PPP5) moved from non-toxic to toxic, despite its true affiliation to the less-toxic class. Finally, MC-RR (in PPP5) moved from the correct prediction of "less toxic" to the wrong prediction of "toxic". Overall, the ML model trained without the MC-LF stereoisomer performed better, producing fewer false-negative predictions.

4 Discussion

From an evolutionary perspective, ser/thr-PPP are a very old family of enzymes, with PPP1 and PPP2A being two of the most slowly evolving enzymes (Cohen, 2004). PPP1 usually works together with a regulatory subunit, while the PPP2A-holoenzyme additionally comprises a scaffold subunit (Shi, 2009). In this regard, PPP2A is similar to PPP4 and PPP6, whose holoenzymes are also comprised of three subunits, i.e., scaffold, regulatory and catalytic (Brewis et al., 1993: Stefansson et al., 2008). Moreover. PPP4 and PPP6 are more closely related to PPP2A than to any other member of the PPP family (Shi, 2009). Thus, it can be assumed that PPP4 and PPP6 behave similarly to MC inhibition as PPP2A. PPP5 is most distant from the other members of the PPP family (Andreeva and Kutuzov, 2001). PPP5 is expressed as a single peptide combining the catalytic domain with a TPR-domain, which interacts with the peptide's c-terminus to act as an auto-inhibitory domain (Yang et al., 2005).

The catalytic subunits of PPP1, PPP2A and PPP5, representing three members of distinct subfamilies of the PPP family, were used for inhibition assays with various MC congeners. In view of the lack of pure human PPP1 available for in vitro testing, PPP1 from rabbit skeletal muscle was used as it displays 100% protein sequence identity to human PPP1 (hPPP1; analyzed with Clustal Omega and Geneious (Biomatters)) and is therefore considered to be equal in structure and enzymatic performance to hPPP1. In contrast, hPPP2A and hPPP5 have only 43% and 37% sequence homology with hPPP1, while hPPP5 and hPPP2A are 40% homologous (analyzed with Clustal Omega). Despite sequence differences, the 3D-structures of the respective catalytic subunits align quite well (see Fig. S6¹), suggestive of similar size and structure restrictions for MC interaction with the hydrophobic groove close to the catalytically active center of the respective PPPs.

Although PPPs have been reported to be inhibited by MC earlier (Fischer et al., 2010; Hastie et al., 2005; Mackintosh et al., 1990; Garibo et al., 2014), the data presented here are unique as they compare the MC-mediated inhibition of the catalytic subunits of three different PPPs in parallel with a hitherto unprecedented number of MC congeners, including synthetically derived structural variants. MC-RR, -LR, -YR, and [β-D-Asp3, Dhb7]-MC-RR presented with similar IC₅₀ for PPP1 and PPP2A (Tab. 1), while PPP2A was \geq 2-fold more sensitive to the more hydrophobic MC-LW, -LA, -LF and the synthetic MC-LF derivatives (MC-LY(Prg), [MSecPh7]-MC-LY(Prg) and [Amba5]-MC-LY(Prg), see Fig. 1. Although ([Amba5]-MC-LY(Prg) displayed an approx. 243-fold higher toxicity towards PPP2A than to PPP1, this was not the case for the other two OPrg-containing congeners ([MSecPh7]-MC-LY(Prg) and [Anda5]-MC-LY(Prg)). Indeed, it has previously been described that MC variants with reduced Adda5-sidechains show a tendency to bind more effectively to PPP2A than to PPP1(Fontanillo et al., 2016).

In principle, the MC congeners' inhibition capacity of PPP5 followed the same trend as observed for PPP1 and PPP2A, albeit being 4 to 200-fold lower. With the exception of [enantio-Adda5]-MC-LF, showing absence of binding to all three PPPs test-



Tab. 3: MC congener toxicity equivalency factors (TEF)

Congener	PPP1		PPP2A		PPP5	PPP5	
	IC ₅₀ (nM)	TEF	IC ₅₀ (nM)	TEF	IC ₅₀ (nM)	TEF	
MC-RR	1.5	0.20	1.6	0.31	11.7	0.44	
MC-LR	0.3	1.00	0.5	1.00	5.1	1.00	
MC-YR	1.3	0.22	n.d.	n.d.	5.1	1.00	
MC-WR	1.2	0.24	1.0	0.50	5.6	0.91	
MC-LA	1.9	0.15	0.7	0.71	6.1	0.84	
MC-LY	0.8	0.36	n.d.	n.d.	4.1	1.24	
MC-LF	2.0	0.15	1.4	0.36	4.7	1.09	
MC-LW	1.2	0.24	0.7	0.66	2.5	2.02	
MC-HilR	0.6	0.49	n.d.	n.d.	4.2	1.20	
MC-HtyR	0.7	0.45	n.d.	n.d.	4.7	1.09	
[β-D-Asp3]-MC-RR	45.0	0.01	n.d.	n.d.	167.1	0.03	
[β-D-Asp3]-MC-LR	0.9	0.34	n.d.	n.d.	10.2	0.50	
[β-D-Asp3, Dhb7]-MC-RR	62.0	0.01	84.3	0.01	877.1	0.01	
MC-LY(Prg)	1.7	0.17	0.4	1.76	1.7	2.95	
[MSecPh7]-MC-LY(Prg)	1.9	0.15	0.9	0.54	18.2	0.28	
[enantio-Adda5]-MC-LF	-	_	_	-	_	-	
[Amba5]-MC-LY(Prg)	520,817	0.000	2,135	0.000	54,063	0.00	
[Anda5]-MC-LY(Prg)	1,724	0.000	n.d.	n.d.	2,420	0.00	

ed (Tab. 1), earlier assumptions regarding size and structure restrictions for MC interaction with the respective PPPs could not be corroborated. Indeed, MC congeners apparently do not bind as tightly to the catalytic subunit of PPP5 as to the catalytic subunits of PPP1 and 2A. Exceptions, surprisingly, were the synthetic MC-LY(Prg) and [Anda5]-MC-LY(Prg), which shared similar inhibitory capabilities in PPP1 and PPP5 (Tab. 1). Asp3 variants of MC-LR appeared to be of comparable toxicity as MC-LR, while in contrast β -D-Asp3 variants of MC-RR were all dramatically less toxic than MC-RR across all PPPs tested.

Although MC-LR is considered to be the most toxic of all congeners (WHO, 2017), this appears to apply only to PPP1 when considering toxicodynamic data. Indeed, MC-LW, MC-LF, MC-LY(Prg) and [MSecPh7]-MC-LY(Prg) presented with comparable inhibitory capabilities as MC-LR in PPP2A. Moreover, in PPP5, MC-WR, -YR, -LY, -LA, -HilR, -HtyR were of comparable toxicity while MC-LF and MC-LY(Prg) were more toxic than MC-LR. The latter suggests that using the toxicity equivalent factors concept (TEF), i.e., all MC congeners equaling in toxicity to MC-LR, would under- and overestimate the potential toxicodynamic capacity present in a given cyanobacterial bloom.

Moreover, the fact that MC congeners have been demonstrated to present with significant differences with regard to OATP transport (Fischer et al., 2005; Feurstein et al., 2011), whereby MC-LR and-RR are transported less efficiently than, e.g., MC-

LA,-LW, or -LF, compounds the problems mentioned with using the TEF as originally proposed by Dietrich and Hoeger (2005). Indeed, in a realistic setting employing a guidance value of 1 µg MC-LR_{equivalent}/L for drinking water (Falconer and Humpage, 2005) and using summary detection methods, e.g., ELISA (Fischer et al., 2001), without concurrent LC-MS/MS confirmation of MC congeners (Puddick et al., 2014) could severely under- or overestimate the toxicity of a MC congener mixture in a given water sample contaminated by a toxic cyanobacterial bloom. Indeed, there are several reports of multiple co-occurring MC congeners in a given cyanobacterial bloom (Falconer et al., 1994; Kleinteich et al., 2018; Puddick et al., 2014), thus demonstrating the reality of having to deal with mixture exposures of different toxicities in a human hazard and risk assessment scenario.

The question then needs to be raised as to how one could deal with the uncertainties of having more than 248 putative MC congeners (Spoof and Catherine, 2017) on one hand, yet absence of relevant toxicity data for the majority of these MC congeners that would allow for appropriate hazard and risk assessment on the other hand. The latter discrepancy is exacerbated by the fact that for most of the 248 putative MC congeners no purified material is available to actually test the MC using *in vitro* toxicokinetic and toxicodynamic assays and thus to provide a minimal dataset that could be of toxicological relevance for humans.



One approach, albeit still limited to the toxicodynamic component of the apical toxicity, is the employed ML approach for roughly categorizing MC congeners into groups of toxic, less toxic, and non-toxic MC, and thus to make preliminary predictions. Although the available data for training of the ML approach was restricted to 18 MC congeners, the number of samples was artificially increased with the oversampling technique SMOTE. This approach introduced uncertainty in the data, which might have caused overlapping data points of different classes and therefore the wrong classification of compounds. Therefore, more MC should be tested, including rare MC variants (e.g., doubly-demethylated congeners), to expand the existing dataset for better model performance and more fine-scaled predictions (i.e., more toxicity classes). Inclusion of other PPP inhibitors, e.g., the structurally related nodularins (Rinehart et al., 1988) and structurally unrelated anabaenopeptins (Spoof et al., 2015), should ensure a better and more sensitive predictive performance. However, irrespective of the potential uncertainties experienced, both models used (Fig. 5, Tab. S4¹) provided 80-90% correct predictions of toxicity class. More importantly however, modulation of the training set improved prediction (Fig. 5B), whereby most of the few wrongly predicted MC were found in a higher toxicity class, thereby overestimating the true toxicity. As overestimation of toxicity would be more precautionary with regard to potential hazard and risk, this caveat of the ML approach was considered acceptable.

Categorizing MC congeners into toxicodynamic classes will not resolve the problem of having to assess the potential hazard of mixtures of different MC congeners present in a given surface water. Additionally, the use of calculated TEF, as attempted by Garibo et al. (2014) with six different MC congeners, or the TEF calculated from the PPP inhibition data in this study (Tab. 3) will not alleviate the problem of lacking toxicokinetic data. Although there have been past efforts to characterize the uptake into human cells via OATPs (Fischer et al., 2005; Fischer et al., 2010; Monks et al., 2007) and current efforts are under way to characterize the efflux of MC congeners from human epithelial cells using human exporter-expressing insect membrane vesicle systems (Kaur et al., 2019), we are still far from actually being able to calculate individual MC congener toxicokinetics or from even considering integrating toxicokinetic and toxicodynamic data into toxicologically based kinetic and dynamic modelling.

Furthermore, the dramatic differences observed for MC congener toxicity in hepatocytes derived from different human patients (Fischer et al., 2010) represent a hurdle that must still be overcome in the future. However, data should be revisited as i) up to 4-fold differences are apparent just in the toxicodynamic (PPP1, 2A and 5) component of the apical toxicity (Tab. 1, 3), ii) rodent and human OATP expression and capacity appear to differ profoundly (Fischer et al., 2005, 2010), and iii) humans differ in OATP type and level of expression, which would mandate that the original safety extrapolations (Dietrich and Hoeger, 2005; WHO, 2017; Falconer and Humpage, 2005) based on rodents be

reconsidered. Prudence would require that the current guidance value of 1.0 MC-LR_{equivalent} μ g/L be lowered by at least a factor 10, thus accommodating recent *in vitro* findings with human cell systems.

In the future, integrated systems toxicology approaches including computational toxicology (Cronin and Dietrich, 2017) and the ML approach presented here in combination with adverse outcome pathways⁴ could provide a much better basis for the hazard and risk assessment of MC in toxic cyanobacterial blooms and thus establish better guidelines regarding the safety of surface waters used for drinking water and recreational purposes.

References

Andreeva, A. V. and Kutuzov, M. A. (2001). PPP family of protein Ser/Thr phosphatases: Two distinct branches? *Mol Biol Evol 18*, 448-452. doi:10.1093/oxfordjournals.molbev. a003823

Asgari, E. and Mofrad, M. R. (2015). Continuous distributed representation of biological sequences for deep proteomics and genomics. *PLoS One 10*, e0141287. doi:10.1371/journal. pone.0141287

Azevedo, S. M. F. O., Carmichael, W. W., Jochimsen, E. M. et al. (2002). Human intoxication by microcystins during renal dialysis treatment in Caruaru-Brazil. *Toxicology 181-182*, 441-446. doi:10.1016/s0300-483x(02)00491-2

Bartram, J. and Chorus, I. (eds.) (1999). *Toxic Cyanobacteria in Water: A Guide to their Public Health Consequences Monitoring and Management.* E & FN Spon. doi:10.1201/9781482295061

Beleites, C., Neugebauer, U., Bocklitz, T. et al. (2013). Sample size planning for classification models. *Anal Chim Acta 760*, 25-33. doi:10.1016/j.aca.2012.11.007

Berry, M. A., Davis, T. W., Cory, R. M. et al. (2017). Cyanobacterial harmful algal blooms are a biological disturbance to Western Lake Erie bacterial communities. *Environ Microbiol 19*, 1149-1162. doi:10.1111/1462-2920.13640

Bowling, L. and Baker, P. D. (1996). Major cyanobacterial bloom in the Barwon-Darling River, Australia, in 1991, and underlying limnological conditions. *Mar Freshw Res* 47, 643-657. doi:10.1071/mf9960643

Brewis, N. D., Street, A. J., Prescott, A. R. et al. (1993). PPX, a novel protein serine/threonine phosphatase localized to centrosomes. *EMBO J 12*, 987-996. doi:10.1002/j.1460-2075.1993. tb05739.x

Chang, N., Vannah, B. and Yang, Y. J. (2014). Comparative sensor fusion between hyperspectral and multispectral satellite sensors for monitoring microcystin distribution in Lake Erie. *IEEE J Sel Top Appl Earth Obs Remote Sens* 7, 2426-2442. doi:10.1109/jstars.2014.2329913

Chawla, N. V., Bowyer, K. W., Hall, L. O. et al. (2002). SMOTE: Synthetic minority over-sampling technique. *J Artif Intell Res 16*, 321-357. doi:10.1613/jair.953

⁴ http://www.oecd.org/chemicalsafety/testing/projects-adverse-outcome-pathways.htm (accessed 28.06.2019)



- Cohen, P. T. (2004). Overview of protein serine/threonine phosphatases. In J. Ariño and D. R. Alexander (eds.), *Protein Phosphatases. Topics in Current Genetics*. Berlin, Heidelberg: Springer. doi:10.1007/978-3-540-40035-6
- Cronin, M. T. D. and Dietrich, D. (2017). Computational toxicology to support predictions of ADMET properties and as a promising tool for 21st century risk assessment. *Comput Toxicol 1*, 1-2. doi:10.1016/j.comtox.2017.02.001
- Dellafiora, L., Dall'Asta, C. and Galaverna, G. (2018). Toxico-dynamics of mycotoxins in the framework of food risk assessment An in silico perspective. *Toxins (Basel)* 10, 52. doi: 10.3390/toxins10020052
- Dietrich, D. and Hoeger, S. (2005). Guidance values for microcystins in water and cyanobacterial supplement products (blue-green algal supplements): A reasonable or misguided approach? *Toxicol Appl Pharmacol* 203, 273-289. doi:10.1016/j. taap.2004.09.005
- EFSA PPR Panel, Ockleford, C., Adriaanse, P. et al. (2018). Scientific opinion on the state of the art of toxicokinetic/toxicodynamic (TKTD) effect models for regulatory risk assessment of pesticides for aquatic organisms. *EFSA J 16*, e05377. doi:10.2903/j.efsa.2018.5377
- Falconer, I. R., Burch, M. D., Steffensen, D. A. et al. (1994). Toxicity of the blue-green alga (cyanobacterium) microcystis aeruginosa in drinking water to growing pigs, as an animal model for human injury and risk assessment. *Environ Toxicol Water Qual* 9, 131-139. doi:10.1002/tox.2530090209
- Falconer, I. R. and Humpage, A. R. (2005). Health risk assessment of cyanobacterial (blue-green algal) toxins in drinking water. *Int J Environ Res Public Health 2*, 43-50. doi:10.3390/ijerph2005010043
- Fawell, J. K., Mitchell, R. E., Everett, D. J. et al. (1999). The toxicity of cyanobacterial toxins in the mouse: I microcystin-LR. *Hum Exp Toxicol 18*, 162-167. doi:10.1177/096032719901800305
- Feurstein, D., Holst, K., Fischer, A. et al. (2009). Oatp-associated uptake and toxicity of microcystins in primary murine whole brain cells. *Toxicol Appl Pharmacol 234*, 247-255. doi:10.1016/j.taap.2008.10.011
- Feurstein, D., Kleinteich, J., Heussner, A. H. et al. (2010). Investigation of microcystin congener-dependent uptake into primary murine neurons. *Environ Health Perspect 118*, 1370. doi:10.1289/ehp.0901289
- Feurstein, D., Stemmer, K., Kleinteich, J. et al. (2011). Microcystin congener- and concentration-dependent induction of murine neuron apoptosis and neurite degeneration. *Toxicol Sci* 124, 424-431. doi:10.1093/toxsci/kfr243
- Fischer, A., Hoeger, S. J., Stemmer, K. et al. (2010). The role of organic anion transporting polypeptides (OATPs/SLCOs) in the toxicity of different microcystin congeners in vitro: A comparison of primary human hepatocytes and OATP-transfected HEK293 cells. *Toxicol Appl Pharmacol* 245, 9-20. doi:10.1016/j.taap.2010.02.006
- Fischer, W. J., Garthwaite, I., Miles, C. O. et al. (2001). Congener-independent immunoassay for microcystins and nod-

- ularins. *Environ Sci Technol 35*, 4849-4856. doi:10.1021/es011182f
- Fischer, W. J., Altheimer, S., Cattori, V. et al. (2005). Organic anion transporting polypeptides expressed in liver and brain mediate uptake of microcystin. *Toxicol Appl Pharmacol* 203, 257-263. doi:10.1016/j.taap.2004.08.012
- Fontanillo, M., Zemskov, I., Häfner, M. et al. (2016). Synthesis of highly selective submicromolar microcystin-based inhibitors of protein phosphatase (PP)2A over PP1. *Angew Chem Int Ed Engl 128*, 14191-14195. doi:10.1002/ange.201606449
- Friedman, J. H. (2001). Greedy function approximation: A gradient boosting machine. *Ann Statist* 29, 1189-1232. doi:10.1214/aos/1013203451
- Garibo, D., Flores, C., Ceto, X. et al. (2014). Inhibition equivalency factors for microcystin variants in recombinant and wild-type protein phosphatase 1 and 2A assays. *Environ Sci Pollut Res Int* 21, 10652-10660. doi:10.1007/s11356-014-3065-7
- Hastie, C. J., Borthwick, E. B., Morrison, L. F. et al. (2005). Inhibition of several protein phosphatases by a non-covalently interacting microcystin and a novel cyanobacterial peptide, nostocyclin. *Biochim Biophys Acta 1726*, 187-193. doi:10.1016/j. bbagen.2005.06.005
- Heresztyn, T. and Nicholson, B. C. (2001). Determination of cyanobacterial hepatotoxins directly in water using a protein phosphatase inhibition assay. *Water Res* 35, 3049-3056. doi:10.1016/s0043-1354(01)00018-5
- Hoeger, S. J., Schmid, D., Blom, J. F. et al. (2007). Analytical and functional characterization of microcystins [Asp3]MC-RR and [Asp3,Dhb7]MC-RR: Consequences for risk assessment? *Environ Sci Technol* 41, 2609-2616. doi:10.1021/es062681p
- Huisman, J., Codd, G. A., Paerl, H. W. et al. (2018). Cyanobacterial blooms. *Nat Rev Microbiol* 16, 471-483. doi:10.1038/s41579-018-0040-1
- Jaeger, S., Fulle, S. and Turk, S. (2018). Mol2vec: Unsupervised machine learning approach with chemical intuition. *J Chem Inf Model* 58, 27-35. doi:10.1021/acs.jcim.7b00616
- Kaur, G., Fahrner, R., Wittmann, V. et al. (2019). Human MRP2 exports MC-LR but not the glutathione conjugate. *Chem Biol Interact 311*, 108761. doi:10.1016/j.cbi.2019.108761
- Kleinteich, J., Wood, S. A., Küpper, F. C. et al. (2012). Temperature-related changes in polar cyanobacterial mat diversity and toxin production. *Nature Clim Change* 2, 356-360. doi:10.1038/nclimate1418
- Kleinteich, J., Puddick, J., Wood, S. A. et al. (2018). Toxic cyanobacteria in Svalbard: Chemical diversity of microcystins detected using a liquid chromatography mass spectrometry precursor ion screening method. *Toxins (Basel)* 10, 147. doi:10.3390/toxins10040147
- Lemaître, G., Nogueira, F. and Aridas, C. K. (2017). Imbalanced-learn: A python toolbox to tackle the curse of imbalanced datasets in machine learning. *J Mach Learn Res* 18, 559-563.
- Mackintosh, C., Beattie, K. A., Klumpp, S. et al. (1990). Cyanobacterial microcystin-LR is a potent and specific inhibitor of protein phosphatase-1 and phosphatase-2A from



- both mammals and higher-plants. FEBS Lett 264, 187-192. doi:10.1016/0014-5793(90)80245-e
- Mikolov, T., Chen, K., Corrado, G. et al. (2013). Efficient estimation of word representations in vector space. arXiv preprint arXiv:1301.3781
- Monks, N. R., Liu, S., Xu, Y. et al. (2007). Potent cytotoxicity of the phosphatase inhibitor microcystin LR and microcystin analogues in OATP1B1- and OATP1B3-expressing HeLa cells. *Mol Cancer Ther* 6, 587-598. doi:10.1158/1535-7163.mct-06-0500
- Pedregosa, F., Varoquaux, G., Gramfort, A. et al. (2011). Scikit-learn: Machine learning in python. *J Mach Learn Res* 12, 2825-2830.
- Pereira, S. R., Vasconcelos, V. M. and Antunes, A. (2011). The phosphoprotein phosphatase family of Ser/Thr phosphatases as principal targets of naturally occurring toxins. *Crit Rev Toxicol* 41, 83-110. doi:10.3109/10408444.2010.515564
- Preece, E. P., Hardy, F. J., Moore, B. C. et al. (2017). A review of microcystin detections in estuarine and marine waters: Environmental implications and human health risk. *Harmful Algae* 61, 31-45. doi:10.1016/j.hal.2016.11.006
- Puddick, J., Prinsep, M. R., Wood, S. A. et al. (2014). High levels of structural diversity observed in microcystins from microcystis CAWBG11 and characterization of six new microcystin congeners. *Mar Drugs* 12, 5372-5395. doi:10.3390/md12115372
- Rinehart, K. L., Harada, K., Namikoshi, M. et al. (1988). Nodularin, microcystin, and the configuration of adda. *J Am Chem Soc 110*, 8557-8558. doi:10.1021/ja00233a049
- Shi, Y. (2009). Serine/threonine phosphatases: Mechanism through structure. Cell 139, 468-484. doi:10.1016/j.cell.2009. 10.006
- Spoof, L., Blaszczyk, A., Meriluoto, J. et al. (2015). Structures and activity of new anabaenopeptins produced by baltic sea cyanobacteria. *Mar Drugs* 14, 8. doi:10.3390/md14010008
- Spoof, L. and Catherine, A. (2017). Appendix 3: Tables of microcystins and nodularin. In J. Meriluoto, L. Spoof and G. Codd (eds.), *Handbook of Cyanobacterial Monitoring and Cyanotoxin Analysis*. John Wiley & Sons, Ltd. doi:10.1002/9781119068761.app3
- Stefansson, B., Ohama, T., Daugherty, A. E. et al. (2008). Protein phosphatase 6 regulatory subunits composed of ankyrin repeat domains. *Biochemistry* 47, 1442-1451. doi:10.1021/bi7022877
- Taranu, Z. E., Gregory-Eaves, I., Steele, R. J. et al. (2017). Predicting microcystin concentrations in lakes and reservoirs at a continental scale: A new framework for modelling an im-

- portant health risk factor. *Global Ecol Biogeogr 26*, 625-637. doi:10.1111/geb.12569
- WHO World Health Organization (2017). *Guidelines for Drinking-Water Quality*. 4th edition. Geneva, Switzerland: World Health Organization. https://www.who.int/water_sanitation_health/publications/drinking-water-quality-guidelines-4-including-1st-addendum/en/
- Wood, S. A., Rueckert, A., Hamilton, D. P. et al. (2011). Switching toxin production on and off: Intermittent microcystin synthesis in a microcystis bloom. *Environ Microbiol Rep 3*, 118-124. doi:10.1111/j.1758-2229.2010.00196.x
- Wu, Z., Ramsundar, B., Feinberg, E. N. et al. (2018). Moleculenet: A benchmark for molecular machine learning. *Chem Sci 9*, 513-530. doi:10.1039/c7sc02664a
- Yang, J., Roe, S. M., Cliff, M. J. et al. (2005). Molecular basis for TPR domain-mediated regulation of protein phosphatase 5. *EMBO J 24*, 1-10. doi:10.1038/sj.emboj.7600496
- Zemskov, I., Altaner, S., Dietrich, D. R. et al. (2017). Total synthesis of microcystin-LF and derivatives thereof. *J Org Chem* 82, 3680-3691. doi:10.1021/acs.joc.7b00175

Conflict of interest

The authors declare that they have no conflicts of interest.

Author contributions

The study was designed by SA, DRD, SJ and FS. Experiments for PPP and PPP2a were performed by SA, while PPP5 experiments were performed by RF under the supervision of SA (Bachelor-thesis RF). Synthetic microcystin variants were produced by IZ under the supervision of VW (PhD Thesis IZ). Machine learning and prediction models were performed by SJ. Manuscript was written by SA, SJ (machine learning part) and DRD. All authors corrected, amended and complemented the manuscript.

Acknowledgments

We gratefully acknowledge the Arthur-und-Aenne-Feindt foundation (Hamburg, Germany), the Konstanz Research School Chemical Biology (KoRS-CB), and CHARM (BadenWürttemberg Wassernetzwerk) for financial support. The UPLC-MS/MS was funded through a large investment grant of the DFG INST 38-537-1.